Estimation of Radioactivity Parameters in Sibaiyaphosphate and Behaviorof Radionuclides during Acidicleachingprocesses

E. S. Abd El-Halim¹, AmanyTahaSroor^{1,*}, NadiaWalley El-Dine¹, Ibrahim E. El-Aassy², Zeinab M. Abdel- GawadShalaby², Naima MAl-kbashy³. ¹ Faculty of Women for Arts, Science and Education, Physics Department, Ain Shams University, Cairo, Egypt

¹ Faculty of Women for Arts, Science and Education, Physics Department, Ain Shams University, Cairo, Egypt ² Nuclear Materials Authority, Cairo, Egypt ³Physics Department, Faculty of Science, Tripoli University, Tripoli, Libya

Corresponding Author: E. S. Abd El-Halim

Abstract: The activity concentrations of natural radionuclides (²³⁸U,²²⁶Ra, ²³⁵U,²³²Th and ⁴⁰K) were measured by using gamma spectrometry in phosphate rock samples collected from phosphate mines located in El-Sibaiya, Nile Valley, Egypt. The average activity concentration of ²³⁸U, ²²⁶Ra, ²³⁵U, ²³²Th and ⁴⁰K (Bq/kg) inphosphate rock were 462.8, 548.08, 21.46, 13.22 and 147.98 respectively. The radiological hazards to the occupational workers caused by y- radiation havebeen determined. Also this study aims to examine the mobility of uranium andits isotopes and daughters during the acid leaching. To study radionuclides transfer from solid material (ore) to the liquid phase (leachate), three samples are prepared for leaching with two different acids .Natural radionuclides were measured by HPGe detector in original samples, pregnant solution and residuals .The results show that the transfer of radionuclides were carried out either physically through a - recoil or chemically through dissolution. The behavior of ²²⁶Ra in the acidic leaching is completely different from from that is not observed in ²³⁴U. The lowest leachability in all acid leaching is present in ²²⁶Ra, ²¹⁴Pb and ²¹⁴Bi. **Key Words:** Phosphaterock, Natural radioactivity, leaching, hazard index.

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I. Introduction

Phosphaterocks contain a relatively high concentration of naturally occurring radioactive materials (NORM) from the uranium and thorium decay series [1]. The average uranium content in phosphate rock is usuallyaroundat 50-200 parts per million(ppm)[1]. Thus the measurement of natural radioactivity in phosphate rocks, which is the basicindicatorof radiological contaminationin the environment, is importantin estimating the present health risks and stablishing the base line data for future monitoring of exploration and in radiation protection[2,3]. The agriculture is the main outlet for phosphate ore as these minerals are important natural sources of fertilizers, in addition to other industrial uses are increasing (detergents, pharmaceutical, metals, food, oil, textiles). For this reason economic exploitation of phosphate ore becomes very important[4].El-Sibaiya cityislocated on he Nile River, the presence of phosphate mines contain an excess of NORM which present a potential risk to the population . In the present work , we measured the concentration of 226Ra, 238U, 232Th, and 40K in phosphate samples collected from some differentsites of El-Sibaiyamines usinga gamma ray spectrometer and calculated the associated absorbed dose rates[2]. In the U-series, the isotope of interest is 230Th, rather than 232Th, which is the parent of another natural decay series. The concentration and occurrence of 230Th depends on its parent 234U, although, chemically, it is reported to associate with refractory elements and the resistant fraction in sediments. 230Th is a good indicator for nuclide migration in the 238U series as it is immobile. The 226Ra/238U ratio is also a good indicator for alteration processes [5] and migration of either 238U or 226Ra. [6]. Uranium isotopes can dissolve from minerals and rocks and enter solution through chemical and physical processes. Physical processes involve α -recoil while chemical processes involve leaching through either acidic or alkaline reagents or any other leaching solution[7]. Thus, observed fractionation between 234U and 238U is generally ascribed to selective leaching, α -recoil transfer of 234Th directly into the aqueous phase, or a combination of the two processes. The principal aim of this work is tostudy radionuclide transfer from solid material (ore) to the liquid phase (leachate)[7]. Acid leaching is known to be predominant process for uranium recovery from ores, usually with sulfuric acid because its relatively low cost [8,9].

Geologic setting

The Sibaiya area is located in the Nile Valley of Egypt at the intersection of latitude 25^005 ' N and longitude 32^0 42 ' E .As shown in figure (1) .



Fig (1) Location of El-Sibaiyaarea [4]

In the region of El-Sibaiya the phosphorite - bearing strata are known as the DuwiFormtion[10]. The DuwiFormation (Phosphate – bearing sediments) is characterized by the presence of three types of sections: a claysection, a sandy- clay section and a sandy section .Three phosphorite - bearing members : lower, middle and upper are recognized in this formation. The phosphate beds could be considered as a probable source of uranium . Where their average content reaches 130 ppm . The Egyptian phosphate beds is due to the accumulation of bones and phosphates derived from the decay of soft parts of organisms and the enrichment of the phosphorite - bearing limestone through the dissolution of calcium carbonates by carbonated waters. There is evidence in the field that processes of metasomatic replacement of calcium carbonate byphosphate took place at Sibaiya [11]. The Duwi Formation starts with thick yellow to yellowish grey, pebbly phosphatic bed (Bed A) with 3 m thickness (Fig2). It istoped with hard silicified phosphate bed (10 cm) which is followed with hard oyster limestone, grey in colour with thin phosphateinterbed (Bed B) of 10 cm thickness and characterized by the presence of shark teeth. Up in the section, the calcareous content is represented by medium hard marl, pale yellow in color with thickness around 1.5 m, intercalated with thin interbred of chert of about 10 cm thicknesses. It overlies by yellow to yellowish grey, fissile shale with a thickness of 25 cm. The phosphatic content is returned again with Bed C that characterized by yellowish grey to brownish grey, medium hard, coated by iron oxides and contains shark teeth; its thickness is 20 cm. This bed forms the top of the Duwi Formation [4].



Fig (2) Lithostratigraphic succession of DuwiFormation exposed in Nile Valley locality [4].

Experimental Procedures

Sampling and samples preparation

For radioactivity measurements , a total of 12 samples each of mass 1 kg , were collected from Sibaiya area is located in the Nile Valley region , Egypt . The collected samples were dried at room temperature for a week crushed , homogenized and sieved through 200 mesh size . The samples were first weighted and placed in polyethylene bottles of 250 cm^3 volume .The bottles were completely sealed for more than one month to allow radioactive equilibrium to be reachedbefore measured by the gamma spectrometer . This stepwas necessary to ensure thatradongas is confined within the volume and the daughters still also remain in the sample . Three phosphate samples (Q7, Q9 at Quseir – Safaga) and (S2 at Sibaiyaarea).Samples are poured from the containers and prepared for leaching experiments using two acids :HCland HNO₃ . Leaching on 150g sample weight under the conditions : solid /liquid ratio 1:3 acid concentration (110 gm/L) , stirring time two hours and at room temperature .Filtration was carried out to separate leachate from residual which is dried at room temperature . The residual and leachate have been weighted and packed will in polyethylene bottles of 250ml volume and stored for more than a month for measuring by HPGe detector. The leachability of the radionuclides was calculated according to the following equation:

About 5 g of each sample was ground to -200 mesh for chemical analyses of major oxides and trace elements content, by the X-ray fluorescence technique (XRF).

Experiment set up

High purity vertical germanium was coupled to a personal computer with a special electronic card to make it equivalent to a multichannel analyzer. The system also contains the usual electronic components of preamplifier, amplifier and power supply. The detector has resolution (FWHM) of 1.85 keV for the 1332.5 keV γ -ray line of ⁶⁰Co. The γ -ray spectrometer energy calibration was performed using ⁶⁰Co, ²²⁶Ra and ²⁴¹Am point sources. The detector was surrounded by a special heavy lead shield of 10 cm thickness with inside dimensions of 28 cm diameter and 40 cm height. The absolute detection efficiency of the HPGe detector was determined by using three well-known reference materials obtained from the International Atomic Energy Agency for U, Th and K activity measurements: RGU-1, RGTh-1 and RGK-1 [12, 13]. The sample containers were placed on top of the detector for counting. The same geometry and size were used for both the samples and the reference materials [14]. The uranium standard (RGU-1) is U-ore diluted with silica with 4940 Bq.kg⁻¹ of ²³⁸U, 228 Bq.kg⁻¹ ¹ of ²³⁵U, a negligible amount of ⁴⁰K (less than 0.63 Bq.kg⁻¹) and some traces of ²³²Th (less than 4 Bq.kg⁻¹). The thorium standard (RGTh-1) is Th-ore diluted with silica having 3250 Bq.Kg⁻¹ of ²³²Th, but containing some ²³⁸U (78 Bq.kg⁻¹) and ⁴⁰K (6.3 Bq.kg⁻¹). The potassium calibration standard (RGK-1) is produced from high purity (99.8 %) potassium sulphate with 14000 Bq.kg⁻¹ of potassium with uranium and thorium contents lower than 0.001 and 0.01 ppm, respectively [13]. The γ -ray transitions used to measure the concentration of the assigned nuclides in the series are the following. ²³⁸U was determined from the gamma rays emitted by its daughter products [**15**]²³⁴Th and ^{234m}Pa activities determined from the 63.3 and 1001 keV photo peaks, respectively, ²¹⁴Bi (609.3, 1120.3, 1238.1, 1377.7 and 1764.5 keV), ²¹⁴Pb (295.1 and 352.0 keV). The specific activity of ²²⁶Ra was measured using the 186.1 keV from its own gamma-ray (after the subtraction of the 185.7 keV of ²³⁵U). The specific activity of ²³²Th was measured using the 338.4, 911.2 and 968.9 keV lines from ²²⁸Ac and 583 keV peak from ²⁰⁸Tl, and ⁴⁰K was measured using 1460.8 keV peak.In order to determine the background contribution due to naturally occurring radionuclides in the environment around the detector, an empty polyethylene beaker of the same 250 cm³ volume was counted with the same geometrical conditions as the sample. The measurement time for both activity and background measurement was 70000 s. The background spectra were used to correct the net gamma- ray peak areas for the studied isotopes.

II. Results and discussion

Activity Concentration

Activity concentration of ²²⁶Ra, ²³⁸U, ²³⁵U, ²³²Th and ⁴⁰K for El-Sibaiya phosphate rockssamples were calculated as illustrated (in Table 1). The activity concentrations of ²³⁸U range (from197.08 to1577.31) Bq/kg, ²²⁶Ra(from 221.4to1599)Bq/kg, ²³⁵Ufrom (9.21 to73.14) Bq/kg , ²³²Th (from 10.38 to 16.44) Bq/kg and⁴⁰K (from96.38to 180.8)Bq/kg. The radioelements worldwide average (W.A.) values are 33 Bq/kg for ²³⁸U, 32 Bq/kg for ²²⁶Ra, 45 Bq/kg for ²³²Th , 412 Bq/kg for ⁴⁰K and 33 Bq/kg for ²³⁵U [**16**] as shown infig (3). The activity concentrations of all studied samples for radium , uranium are higher than the permissible level, but the activity concentrations of most studied samples for ²³⁵Uare lower than the permissible level except samples (S5, S11) are higher than the permissible level.

Table (1) : The activity concentrations in (Bq/kg) of 238 U , 226 Ra , 235 U, 232 Th, 40 K and 226 Ra/ 238 U for phosphate samples at Sibayia area .

Samples	²³⁸ U	²²⁶ Ra	²³⁵ U	²³² Th	⁴⁰ k	²²⁶ Ra/ ²³⁸ U
S1	444.14	533.5	20.76	16	168.8	1.2
S2	442.76	673.82	21.03	10.97	180.8**	1.52
S 3	355.52	548.18	16.32	15.35	170.3	1.54**
S4	309.57	365.7	14.12	12.11	144	1.1
S5	874.46	969.9	40.3	14.11	125.2	1.1
S6	326.1	426.9	15.05	11.12	155.5	1.3
S7	267.85	364.2	12.49	13.32	155.6	1.3
S 8	256	294.2	11.78	12.08	134.3	1.1
S9	197.08*	221.4*	9.21*	10.83	158	1.1
S10	299.96	346.9	13.99	10.38*	140.1	1.1
S11	1577.31**	1599**	73.14**	16.44**	96.38*	1*



Fig(3)The activity concentrations for ²³⁸U, ²²⁶Ra, ²³⁵U, ²³²Th, ⁴⁰K in the phosphate samples in comparison with the permissible level (p.l.).

In this work the activity concentration of 238 U and 232 Th in phosphate samples (in ppm) and activity ratio 232 Th/ 238 U were calculated and illustrated in (table 2).The concentrations of 238 U range between 15.89 and 127.2ppm, and 232 Th between 2.55and 4.04 ppm, while the 232 Th/ 238 U ratios range between 0.03and 0.24 which is lower than the Clark's value (3.5), which indicates that these phosohatesare enriched in uranium.

	²³⁸ U	²³⁵ U	²³² Th	⁴⁰ K	
Sample	(ppm)	(ppm)	(ppm)	%	²³² Th/ ²³⁸ U
S1	35.81	0.25	3.94	0.53	0.11
S2	35.7	0.26	2.7	0.57**	0.07
S3	28.67	0.2	3.78	0.54	0.13
S4	24.96	0.17	2.98	0.46	0.11
S 5	70.52	0.5	3.47	0.4	0.04
S6	26.29	0.18	2.73	0.49	0.1
S7	21.6	0.15	3.28	0.49	0.15
S8	20.64	0.14	2.97	0.42	0.14
S 9	15.89 *	0.11 *	2.66	0.5	0.16
S10	24.19	0.17	2.55*	0.44	0.1
S11	127.2**	0.91**	4.04 **	0.3*	0.03*
S12	16.36	0.11	3.94	0.46	0.24**
* The lowest value			** T	he highest valu	e

Radiologicalhazard indices

Radium Equivalent (Ra_{eq}):

Radium equivalent (Ra_{ea}) index in Bq/kg is a widely used radiological hazard index. It is a convenient index to compare the specific activities of samples containing different concentrations of 226 Ra, 232 Th and 40 K. It was defined on the assumption that 10Bq/kg of 226 Ra, 7Bq/kg of 232 Th and 130Bq/kg of 40 K produce the same gamma dose rate. It was calculated as follows.

 $Ra_{eq} = C_{Ra} + 1.43 C_{Th} + 0.077 C_{k}$ (1) Where C_{Ra} , C_{Th} and C_{k} are the activity concentrations of ²²⁶Ra, ²³²Th and ⁴⁰K in Bq/kg, respectively. Ra_{eq} was estimated for the collected samples and are given in Table (3), were estimated higher than the recommended maximum value of 370 Bq kg⁻¹ except (S8, S9, S12) lower than the recommended maximum value[17].

External Hazard Index (Hex) :

The external hazard index (Hex) represents the external radiation exposure associated with gamma irradiation from radionuclides of concern. The value of H_{ex} should not exceed the maximum acceptable value of one in order to keep the hazard insignificant. The external hazard index (H_{ex}) is defined by equation [18]:

$$H_{ex} = (C_{Ra}/370 + C_{Th}/259 + C_{k}/4810) \le 1 \qquad \dots \qquad (2)$$

Where C_{Ra} , C_{Th} and C_K are the concentration in(Bq Kg⁻¹) of radium, thorium and potassium respectively. The values of external hazard Indices for all samples are found to be more than 1 except samples (S8, S9, S12) lower than the recommended maximum value and sample S10 is found at unity (shown Table (3), Fig (4).

Radiation Level Index (I_{γ}) :

This index can be used to estimate the level of γ -radiation hazard associated with the natural radionuclides in the samples; it is given by the equation

 $I_{\gamma} = C_{Ra'} (150 + C_{Th} / 100 + C_K / 1500 \qquad (3)$ where C_{Ra} , C_{Th} and C_K are the activity concentration of ²²⁶Ra, ²³²Th, ⁴⁰K in Bq/kg, respectively. The value of these indices must be less than unity in order to keep the radiation hazard insignificant, the values of **Representative Level**

Index (I_{γ}) for all samples are found to be more than 1 given in Table (3), Fig (4).

Absorbed Dose Rate D:

The absorbed gamma dose rates in air at 1 m above the ground surface for the uniform distribution of radionuclides (238 U, 232 Th and 40 K) were calculated by using equation :

 $D = 0.427 * C_{U} + 0.662 * C_{Th} + 0.043 * C_{K} (nGy/h) \qquad (4)$ where C_{U} , C_{Th} and C_{K} are the activity concentration of ²³⁸U, ²³²Th and ⁴⁰K in Bq/kg, respectively. The range of absorbed dose rate in the samples due to natural radionuclidesis (98.09–688.5) nGy h^{-1} given in Table (3). The values of the Dose rate calculated during present study are found higher than the permissible level of 59 nGy/h[19].

The annual outdoor effective dose (E_{out}) is estimated from the outdoor external dose rate (D_{out}), time of stay in the outdoor or occupancy factor (OF = 20 % of 8760 h in a year) and the conversion factor (CF = 0.7 Sv.Gy⁻¹) to convert the absorbed dose in air to effective dose. During the present study, the Eout was calculated using the following equation from[16]:

 $E_{out} = D_{out} (nGy h^{-1}) * 0.2 * 8760 h * 0.7 (Sv * Gy^{-1})$(5)

The outdoor effective dose rate E_{out} range from (120.26to 844.1mSv/y), the values of the all studied samples listed (in table 3) are found higher than the world's average 0.07 mSv/y. Excess lifetime cancer risk (ELCR) :

The value of annual effective dose excess lifetime cancer risk (ELCR) was calculated by using the equation : $ELCR_{out} = E_{out*} LE * RF$

Where E_{out} is the annual effective dose, LE life expectancy (66 years) and RF (Sv⁻¹) is risk factor per Sievert, which is 0.05. The values of ELCR_{out} range between (396.87x10⁻³- 2785.55x10⁻³) in the phosphate samples Table (3). All values are higher than the permissible level 0.29×10^{-3} [20].

Table (3): The values of radium equivalent (Bq/kg), radioactivity level index, external hazard, the outdoor absorbed rate (nGy/h), outdoor annual effective dose (mSv/y) and outdoor Excess lifetime cancer risk at Sibayia area .

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Sample	Ra _{eq} (Bq/kg)	Ι _γ	H _{ex}	D _(out) (nGy/h)	$\frac{E_{eff(out)}}{(mSv/y)}$	ELCR _(out) x10 ⁻³
S1	569.37	3.82	1.53	207.46	254.35	839.36
S2	703.42	4.72	1.9	204.07	250.19	825.63
S 3	583.24	3.92	1.57	169.26	207.51	684.79
S4	394.1	2.65	1.06	146.37	179.45	592.18
S5	999.71	6.69	2.7	388.09	475.79	1570.13
S6	454.77	3.06	1.22	153.27	187.9	620.1
S7	395.22	2.66	1.06	129.85	159.2	525.36
S8	321.81	2.17	0.86	123.05	150.87	497.87
S9	249.05*	1.68*	0.67*	98.09*	120.26*	396.87*
S10	372.53	2.5	1.0	140.95	172.81	570.28
S11	1629.93**	10.88**	4.4**	688.5**	844.1**	2785.55**
S12	267.48	1.81	0.72	103.51	126.9	418.8
P.L	370 Ba/kg	unity	unity	59 nGvh ⁻¹	0.07 mSvv ⁻¹	0.29*10^-3

*The lowest value

****The highest value**



Fig (4) The values of radium equivalent (Bq/kg), radioactivity level index, external hazard, the outdoor absorbed rate (nGy/h), outdoor annual effective dose (mSv/y) and outdoor Excess lifetime cancer risk at Sibayia area .

LeachingStudies

solutions).							
Sample Q7 leaching by HCI							
Radionuclide	Original (Bq/Kg)	Residual (Bq/Kg)	Solution (Bq/l)	Leachability %	Summation (Residual +solution) %		
²³⁸ Useries							
²³⁴ Th	811.29 ± 24.3	687.65±20.6	173.7 ± 5.2	21.41	106.17		
^{234m} Pa	797.94 ± 23.9	702.4 ±21	132.96 ± 3.9	16.66	104.69		
Average	804.62±24.1	695.02±20.8	153.33 ± 4.5	19.05	105.43		
²³⁴ U	696.15±20.8	587.29±17.6	174.27 ±5.2	25.03	109.39		
²³⁰ Th	657.36 ±19.7	696.57±20.8	22.62 ± 0.6	3.44	109.4		
²²⁶ Ra subseries							
²²⁶ Ra	964.81±28.9	1035±31	101±3	10.46	117.74		
²¹⁴ Pb	682.75±20.4	457.23±13.7	21.35 ± 0.6	3.12	70.09		
²¹⁴ Bi	701.93 ± 21	616.59±18.4	78.61 ± 2.3	11.19	99.04		
²³⁵ U	37.63±1.1	36.08±1	7.16 ± 0.2	19.05	114.93		
²³² Th series							
²²⁸ Ac	12.6 ± 0.3	9.49±0.2	3.67 ± 0.1		104.44		
²⁰⁸ Tl	14.79±0.4	13.09 ± 0.3	4.42 ± 0.1		118.31		
Average	13.69 ± 0.4	11.29±0.3	4.05 ± 0.1		112.05		
⁴⁰ K	116.4 ± 3.4	$1174{\pm}35$	253.16±7.5		1226.08		
²³⁸ U/ ²³⁵ U	21.38	19.26	21.38				
²³⁴ U/ ²³⁵ U	18.5	16.27	24.3				
²³⁴ U/ ²³⁸ U	0.86	0.84	1.13				
²²⁶ Ra/ ²³⁸ U	1.19	1.48	0.65				

 Table (4) Activity concentrations of radionuclides in original samples, residuals and leachates (pregnant solutions).

Table (5) Continued

Sample Q7 leaching by HNO ₃							
Radionuclide	Original (Bq/Kg)	Residual (Bq/Kg)	Solution (Bq/l)	Leachability %	Summation (Residual +solution) %		
²³⁸ Useries							
²³⁴ Th	811.29±24.3	665.18±19.9	155.59 ± 4.6	19.17	101.16		
^{234m} Pa	797.94±23.9	702.4 ±21	141.43±4.2	17.72	105.75		
Average	804.62±24.1	683.79±20.5	148.51±4.4	18.45	103.44		
²³⁴ U	696.15±20.8	537.71±16.1	220.36 ±6.6	31.65	108.89		
²³⁰ Th	657.36±19.7	409.75±12.2	76.23±2.2	11.59	73.93		
²²⁶ Ra subseries							
²²⁶ Ra	964.81±28.9	931.59±27.9	197±5.9	20.41 8	116.97		
²¹⁴ Pb	682.75±20.4	637.8±19.1	17.17 ± 0.5	2.51	95.93		
²¹⁴ Bi	701.93±21	743±22.2	76.02±2.2	10.83	116.68		
²³⁵ U	37.63±1.1	31.03±0.9	6.98 ± 0.2	18.56	101.02		
²³² Th series							
²²⁸ Ac	12.6±0.3	11.79±0.3	2.65±0.07		114.63		
²⁰⁸ Tl	14.79 ± 0.4	13.09 ±0.3	4.06±0.1		115.89		
Average	13.69±0.4	12.44 ± 0.3	3.35 ± 0.1		115.35		
⁴⁰ K	116.4±3.4	950.59±28.5	358.78±10.7		1124.89		
$^{238}U/^{235}U$	21.38	22.03	21.25				
²³⁴ U/ ²³⁵ U	18.5	17.33	31.54				
²³⁴ U/ ²³⁸ U	0.86	0.78	1.48				
²²⁶ Ra/ ²³⁸ U	1.19	1.36	1.32				

Table (6) Continued							
Sample Q9 leaching by HCl							
Radionuclide	Original (Bq/Kg)	Residual (Bq/Kg)	Solution (Bq/l)	Leachability %	Summation (Residual +solution)		
²³⁸ Useries					/0		
²³⁴ Th	1135.47 ±34	1101.6 ±33	154.22±4.6	13.58	110.59		
^{234m} Pa	1108.48±33.2	967.18±29	151.74±4.5	13.68	100.94		
Average	1121.97±33.6	1034.39 ±31	152.98±4.5	13.63	105.83		
²³⁴ U	1008.11±30.2	611.31±18.3	292.34±8.7	28.99	89.63		
²³⁰ Th	544.87±16.3	677.06±20.3	40.59±1.2	7.45	131.71		
²²⁶ Ra subseries							
²²⁶ Ra	2076 ± 62.2	2097.25±62.9	77.62 ± 2.3	3.73	104.76		
²¹⁴ Pb	968.43±29	497±14.9	72.2±2.1	7.45	58.77		
²¹⁴ Bi	1200±36	1171±35.1	105.2±3.1	8.76	106.35		
²³⁵ U	51.39±1.5	53.97 ±1.6	7.1 ± 0.2	13.81	118.84		
²³² Th series							
²²⁸ Ac	10.29 ± 0.3	7.18±0.2	3.97±0.1		108.51		
²⁰⁸ Tl	6.94 ±0.2	4.5±0.1	3.66 ±0.1		117.64		
Average	8.61 ± 0.2	5.84±0.1	3.82±0.1		112.22		
⁴⁰ K	118±3.5	1755.53 ± 52.6	284±8.5		1728.41		
$^{238}U/^{235}U$	21.83	19.16	21.53				
$^{234}U/^{235}U$	19.61	11.32	41.14				
$^{234}U/^{238}U$	0.89	0.59	1.91				
²²⁶ Ra/ ²³⁸ U	1.85	2.02	0.5				

Table (7) Continued

Sample Q9 leaching by HNO ₃						
Radionuclide	Original (Bq/Kg)	Residual (Bq/Kg)	Solution (Bq/l)	Leachability %	Summation (Residual +solution) %	
²³⁸ Useries						
²³⁴ Th	1135.47±34	934.85 ± 28	144.92±4.3	12.76	95.09	
^{234m} Pa	1108.48±33.2	702.4 ± 21	170.28±5.1	15.36	78.72	
Average	1121.97±33.6	818.62±24.5	157.6 ±4.7	14.04	87.01	
²³⁴ U	1008.11±30.2	943.32 ± 28.2	218.16±6.5	21.64	115.21	
²³⁰ Th	544.87±16.3	532.67±15.9	36.94±1.1	6.77	104.54	
²²⁶ Ra subseries						
²²⁶ Ra	2076±62.2	2100±63	208.1±6.2	10.02	111.18	
²¹⁴ Pb	968.43±29	1030±30.9	21.87±0.6	2.25	108.61	
²¹⁴ Bi	1200±36	1370±41.1	67.29±2	5.6	119.77	
²³⁵ U	51.39±1.5	37.95±1.1	7.21±0.2	14.02	87.89	
²³² Th series						
²²⁸ Ac	10.29±0.3	8.82±0.2	3.28±0.09		117.59	
²⁰⁸ Tl	6.94 ±0.2	4.9±0.1	2.6±0.07		108.05	
Average	8.61 ±0.2	6.86±0.2	2.94±0.08		113.8	
⁴⁰ K	118±3.5	3668±110	508.3±15.2		3539.38	
²³⁸ U/ ²³⁵ U	21.83	21.56	21.85			
²³⁴ U/ ²³⁵ U	19.61	24.85	30.24			
²³⁴ U/ ²³⁸ U	0.89	1.15	1.38			
²²⁶ Ra/ ²³⁸ U	1.85	2.56	1.32			

Table (8) Continued							
Sample S2 leaching by HCl							
Radionuclide	Original (Bq/Kg)	Residual (Bq/Kg)	Solution (Bq/l)	Leachability %	Summation (Residual +solution) %		
²³⁸ Useries							
²³⁴ Th	444.31±13.3	471.92 ±14.1	143.22±4.2	32.23	138.45		
^{234m} Pa	441.21±13.2	301.02 ± 9	146.72±4.4	33.51	102.27		
Average	442.76±13.2	386.47±11.5	144.97 ± 4.3	32.87	120.49		
²³⁴ U	447.35 ±13.4	410.77±12.3	213.95±6.4	47.82	139.64		
²³⁰ Th	545.41±16.3	409.75±12.2	30.99±0.9	5.68	80.8		
²²⁶ Ra subseries							
²²⁶ Ra	673.82±20.2	665.42±19.9	38.77 ±1.1	5.75	104.5		
²¹⁴ Pb	424.43±12.7	346.5±10.3	6.82±0.2	1.6	83.24		
²¹⁴ Bi	439.53±13.1	405.67±12.1	40.66±1.2	9.25	101.54		
²³⁵ U	21.03±0.6	19.87±0.5	7.31 ± 0.2	34.76	129.27		
²³² Th series							
²²⁸ Ac	13.81 ±0.4	9.41 ±0.2	1.36 ±0.04		78.06		
²⁰⁸ Tl	8.11 ±0.2	14.72±0.4	2.39 ±0.07		210.86		
Average	10.96±0.3	12.07 ± 0.3	1.88 ±0.05		127.21		
⁴⁰ K	180.81±5.4	1062.43 ± 31.8	$\textbf{325.94} \pm \textbf{9.7}$		767.83		
²³⁸ U/ ²³⁵ U	21.05	19.44	19.82				
²³⁴ U/ ²³⁵ U	21.26	20.76	29.25				
²³⁴ U/ ²³⁸ U	1.01	1.06	1.47				
²²⁶ Ra/ ²³⁸ U	1.52	1.72	0.26				

Table (9) Continued

Sample S2 leaching by HNO ₃							
Radionuclide	Original (Bq/Kg)	Residual (Bq/Kg)	Solution (Bq/l)	Leachability %	Summation (Residual +solution) %		
²³⁸ Useries							
²³⁴ Th	444.31±13.3	369.12±11	173.03 ± 5.1	38.94	122.02		
^{234m} Pa	441.21±13.2	235.45±7	155.63±4.6	35.55	89.33		
Average	442.76 ± 13.2	302.29 ±9	164.33 ±4.9	37.26	105.8		
²³⁴ U	447.35 ±13.4	573.22±17.1	198.87 ±5.9	44.45	172.59		
²³⁰ Th	545.41±16.3	961.5 ±28.8	38.58 ± 1.1	7.07	183.36		
²²⁶ Ra subseries							
²²⁶ Ra	673.82 ± 20.2	665.06 ±19.9	85.46 ±2.5	12.68	111.38		
²¹⁴ Pb	424.43 ± 12.7	613.29±18.3	13.97 ± 0.4	3.29	147.79		
²¹⁴ Bi	439.53 ± 13.1	562 ±16.8	13.53±0.4	3.07	130.94		
²³⁵ U	21.03±0.6	13.83 ±0.4	7.96 ±0.2	37.86	103.61		
²³² Th series							
²²⁸ Ac	13.81 ± 0.4	11.05±0.3	1.45 ±0.04		90.5		
²⁰⁸ Tl	8.11 ±0.2	11.51±0.3	2.44 ±0.07		172.02		
Average	10.96 ± 0.3	11.28±0.3	1.95±0.05		120.67		
⁴⁰ K	180.81±5.4	1115.31±33.4	321.52±9.6		794.63		
²³⁸ U/ ²³⁵ U	21.05	21.85	20.63				
²³⁴ U/ ²³⁵ U	21.26	41.44	24.88				
$^{234}U/^{238}U$	1.01	1.89	1.2				
²²⁶ Ra/ ²³⁸ U	1.52	2.2	0.52				

Radiometric Measuremrnts

The results of γ -detection of radionuclides in the original samples (Bq/kg), residuals (Bq/kg) and leachates (Bq/l) are collected in tables (4-9). The activity concentration of ²³⁸U in the original samples, residuals and pregnant solutions for all samples are higher than the typical world average value of **33**Bq/kg [**16**]. The behavior of each radionuclide in the samples shows that the ²²⁶Ra, ²¹⁴Pb and ²¹⁴Bi activity concentrations are higher in the residual than that in the original for all samples exceptsamples Q7, Q9 leaching by HCl.The activity concentration of ²²⁶Ra in the original samples, residuals and pregnant solutions are higher than the typical world average value of **32** Bq/kg,[**16**]. For ²³²Th, all the original samples ,residuals and pregnant solutions are lower than the permissible level **45**Bq/kg. For all samples⁴⁰K, the original samples are lower than the permissible level **412**Bq/kg while the residuals are higher .

Geochemical Characterization of Samples

Three phosphate samples ($\overline{Q7}$, Q9 at Quseir – Safaga) and (S2 at Sibaiyaarea)were prepared and analyzed chemically for determination of major oxides order of a (%). The distributions of these components are of significant values of SiO₂ , Al₂O₃ , Fe₂O₃, TiO₂ , CaO , MgO , Na₂O , K₂O , P₂O₅ and L.O.I(loss on ignition), table (10) . From this table, the sample (S2) has high level of SiO₂ ,CaOandL.O.I contents are 45.57, 23.40 and 12.50 % respectively . The sample (Q7) has high level of CaO and L.O.I which contents are 44.00 , 32.02% respectively . The sample (Q9) has high level of CaO , P₂O₅ and L.O.I.Which contents are 50.40 , 20.50 and 19.10% respectively .

Locality	Sibayia	Quseir – Safaga		
Major oxides	S2 %	Q7 %	Q9 %	
SiO ₂	45.57	3.06	1.78	
Al ₂ O ₃	1.9	2.11	1.96	
Fe ₂ O ₃	2.69	1.59	2.47	
Ti ₂ O	0.41	0.02	0.08	
CaO	23.40	44.00	50.40	
MgO	1.56	4.22	1.89	
Na ₂ O	1.50	1.65	0.99	
K ₂ O	0.79	0.26	0.04	
P ₂ O ₅	8.61	11.20	20.50	
L.O.I	12.50	32.02	19.10	
Total	99.03	100.13	99.21	

 Table (10) Major elements in Phosphate samples :

The concentrations of trace elements in phosphate samples at Quseir – Safaga Area (Q6, Q7, Q9) and phosphate samples at Sibaiya area (S2,S4,S8) were measured and listed in Table (11). The concentrations of Cr (Chromium) vary from 133 to 211 ppm, Cu (Copper) vary from 17 to 37 ppm, Ni (Nickel) range from 13 to 66 ppm, Zn (Zinc) range from 133 to 411 ppm, Zr (Zirconium) vary from 12 to 28 ppm, Sr (Strontium) varyfrom 570 to 1177 ppm, Y(Yttrium) varyfrom 15to 39 ppm and Pb(Lead) range from 2 to 9 ppm, while two elements V(Vanadium) and Ba (Barium) (the lowest values of phosphate samples at Quseir – Safagaarea(Q6,Q7,Q9) and the highest values of phosphate samples at Sibaiya area(S2,S4,S8).

Locality	Quseir – Safaga			Sibayia		
Trace Element	Q6 (ppm)	Q7 (ppm)	Q9 (ppm)	S2 (ppm)	S4 (ppm)	S8 (ppm)
Cr	188	211	162	147	133	178
Cu	37	23	17	25	23	20
Ni	22	13	32	66	18	16
Zn	261	198	311	411	155	133
Zr	18	12	15	21	17	28
Sr	972	570	828	1177	857	1084
Y	39	26	24	18	15	18
V	1.5	1.5	1.5	1996	2328	2100
Pb	6	2	4	9	4	3
Ba	1.5	1.5	1.5	852	990	917

Table (11) The concentrations of the trace elements in (ppm) for the Phosphate samples .

III. Conclusions

It is noticed that, there is a difference in γ -activity between the summation of activities of residual and pregnant solution with the activity of the original sample. This phenomenon is varied in magnitude within the different radionuclides. The type of sample plays its role in these variations. But the activities of residual and pregnant solution in all samples for ⁴⁰K are much larger than the original , probably the reason is due tochemical reactions which hapend during leaching process .The behavior of radionuclides (²³⁸U, ²³⁵U, ²³⁴U, ²²⁶Ra, ²¹⁴Pb, ²¹⁴Bi, ²³²Th and ⁴⁰K) was studied during the acid leaching process, the ²³⁵U has nearly the same leachability as ²³⁸U.The uranium isotopes are leached to the same extent but that the same is not observed between ²³⁴U and ²³⁸U due to the effect of α -recoil phenomenon.The lowestleachability in all the samples is present in ²¹⁴Pb between(1.6% - 7.45%). Theleachability (leaching efficiency %) present in sample (S2 at Sibaiyain Nile Valley)by HCl and HNO₃(Tables8-9)is more than theleachabilitypresents in samples(Q7,Q9 at Quseir – Safaga)due to the action of carbonate contents in the last two samples.

References

- [1]. International Atomic Energy Agency(IAEA).Extent of Environmental Contamination by Naturally Occurring Radioactive Material (NORM) and Tech- nological Options for Mitigation 2003, Technical:208
- [2]. HARB S, AHMED N, BADAWY W and SAADN ..(2016):"Estimation of radioecological parameters of soil samples from a phosphtic area "Nu clear Technology& Radiation Protection: , Vol. 31, No. 2, pp. 165-172
- [3]. White, G. J., Rood, A. S., Radon Emanation from NORM-Contaminated Pipe Scale and Soil at Petro -leum Industry Sites, J.

- [4] [4] Issa Shams A. M. Mostafa A. M. A. LotfyAbd El-Salam M., (2015) "Radiological impacts of natural radioactivity in phosphate rocks from El- Sibaiya and Red Sea coast mines, Egypt " J RadioanalNuclChem 303:53–61
- [5]. I.E. El Aassy, M.M. El Galy, A.A. Nada, M.G. El- Feky, Th.M. Abd El Maksoud, Sh.M. Talaat, and E.M. Ibrahim, "Effect of alteration processes on the distribution of radionuclides in uraniferous sedimentary rocks and their environmental impact, southwestern Sinai, Egypt," Journal of Radioanalytical and Nuclear Chemistry, vol. 289, pp. 173–184, 2011.
 [6]. El AassyI E., El Feky M G, El Kasaby M. A, Ibrahim E M, Sewefi S, AttiaR M " (2017) Behavior of Radionuclides during
- [6]. El Aassyl E., El Feky M G, El Kasaby M. A, Ibrahim E M, Sewefi S, AttiaR M " (2017) Behavior of Radionuclides during Acidic Leaching Processes of Different Rock Materials, Allouga Locality, Southwestern Sinai, Egypt " International Journal of Scientific & Engineering Research, Volume 8, Issue 1, ISSN 2229-5518
- [7]. M.B. Andersen, Y. Erel, and B. Bourdon, "Experimental evidence for ²³⁴U-²³⁸U fractionation during granite weathering with implication for ²³⁴U/²³⁸U in natural waters," Geochim.CosmochimActa, vol. 73, pp. 4124-4141, 2009.
- [8]. Umanskii, A.B. and Klyushnikov A.M., (2012): "Bioleaching of low grade uranium ore containing pyrite using A. ferrooxidansand A. thiooxidans". Journal of Radio analytical and Nuclear Chemistry.DOI 10.1007/s10967- 012-1816-9.
- [9]. Abd El-Halim E.S., Sroor .A, El-Bahi S.M., El-Aassyl E., El Sheikh E. M , Musa K.M. (2017) "Factors Controlling Radionuclides Migration within Different Media " IOSR Journal of Applied Physics (IOSR-JAP) e-ISSN: 2278-4861.Volume 9, Issue 6 Ver.III, PP 34-41
- [10]. WilliamBoswortha, Philippe Huchonb and Ken McClayc, (2005). "The Red Sea and Gulf of Aden Basins", Journal of A frican Earth Sciences, Volume 43, Issues 1-3, Pages 334-378. Phanerozoic Evolution of Africa.
- [11]. GamalS.Awadalla ,(2010) " Geochemistry and microprobe investigations of Abu Tartur REE bearing phosphorite, Western Desert, Egypt ", Journal of African Earth Sciences, Volume 57, Issue 5, July 2010, Pages 431 443.
- [12]. Preparation and certification of IAEA gamma-ray spectrometry reference materials. RGU-1, RGTh-1 and RGK-
- 1.IAEA/RL/148.International Atomic Energy Agency, 1987.54 p.
- [13]. R.M. Anjos et al. Natural radionuclide distribution in Brazilian commercial granites.Radiat.Meas. 39 (2005) 245.
 [14]. S. Turhan, L. Gündüz. Determination of specific activity of ²²⁶Ra, ²³²Th and ⁴⁰K for assessment of radiation hazards from Turkish
- [14]. S. Turhan, L. Gündüz. Determination of specific activity of ²²⁶Ra, ²³²Th and ⁴⁰K for assessment of radiation hazards from Turkish pumice samples. J. Environ. Radioact.99 (2008) 332.
- [15]. R.A. Sutherland, E. de Jong. Statistical analysis of gamma-emitting radionuclide concentrations for three fields in southern Saskatchewan, Canada.Health Phys 58 (1990) 417.
- [16]. UNSCEAR, 2010. Sources and Effects of Ionizing Radiation. Report to General Assembly with Scientific Annexes (New York, United Nations Scientific Committee on the Effect of Atomic Radiation).
- [17]. AbabnehZ .Q, El-Omari ,H. Rasheed, M. El-Najjar T. and A. Ababneh, "Assessment of gamma emitting radionuclides in sediment cores from the Gulf of Aqaba", Red Sea.RadiatProtDosim., vol. 141, (2010), pp. 289-298.
- [18]. Jankovic, M. Todorovic, D and Savanovic, M. "Radioactivity measurements in soil samples collected in the Republic of Srpska", Radiation Measurements, vol. 43, (2008), pp. 1448-1452.
- [19]. Abd El-Halim E. S., Walley El-Dine N, El-Bahi1 S.M, El-Aassy I.E, El-Sheikh E.M, Al-AbrdiA.M " Excessive lifetime cancer risk and natural radioactivity measurements of granite and sedimentary rock samples " ISSN 1818-331X (2017) T. 18 № 4.
- [20]. Qureshi A.A. et al. Evaluation of excessive lifetime cancer risk due to natural radioactivity in the rivers sediments of Northern Pakistan. Journal of Radiation Research and Applied Science 7 (2014) 438

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Environ. Radioact., 54 (2001),3, pp. 401-413